



Sensitivity of global
freshwater fluxes

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Sensitivity of simulated global-scale freshwater fluxes and storages to input data, hydrological model structure, human water use and calibration

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Abstract

Global-scale assessments of freshwater fluxes and storages by hydrological models under historic climate conditions are subject to a variety of uncertainties. Using the global hydrological model WaterGAP 2.2, we investigated the sensitivity of simulated freshwater fluxes and water storage variations to five major sources of uncertainty: climate forcing, land cover input, model structure, consideration of human water use and calibration (or no calibration). In a modelling experiment, five variants of the standard version of WaterGAP 2.2 were generated that differed from the standard version only regarding the investigated source of uncertainty. Sensitivity was analyzed by comparing water fluxes and water storage variations computed by the variants to those of the standard version, considering both global averages and grid cell values for the time period 1971–2000. The basin-specific calibration approach for WaterGAP, which forces simulated mean annual river discharge to be equal to observed values at 1319 gauging stations (representing 54 % of global land area except Antarctica and Greenland), has the highest effect on modelled water fluxes and leads to the best fit of modelled to observed monthly and seasonal river discharge. Alternative state-of-the-art climate forcings rank second regarding the impact on grid cell specific fluxes and water storage variations, and their impact is ubiquitous and stronger than that of alternative land cover inputs. The diverse model refinements during the last decade lead to an improved fit to observed discharge, and affect globally averaged fluxes and storage values (the latter mainly due to modelling of groundwater depletion) but only affect a relatively small number of grid cells. Considering human water use is important for the global water storage trend (in particular in the groundwater compartment) but impacts on water fluxes are rather local and only important where water use is high. The best fit to observed time series of monthly river discharge (Nash–Sutcliffe criterion) or discharge seasonality is obtained with the standard WaterGAP 2.2 model version which is calibrated and driven by a sequence of two time series of daily observation-based climate forcings, WFD/WFDEI. Discharge computed by a calibrated model version using

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monthly CRU 3.2 and GPCC v6 climate input reduced the fit to observed discharge for most stations. Taking into account the investigated uncertainties of climate and land cover data, we estimate that the global 1971–2000 discharge into oceans and inland sinks is between 40 000 and 42 000 km³ yr⁻¹. The range is mainly due differences in precipitation data that affect discharge in uncalibrated river basins. Actual evapotranspiration, with approximately 70 000 km³ yr⁻¹, is rather unaffected by climate and land cover in global sum but differs spatially. Human water use is calculated to reduce river discharge by approximately 1000 km³ yr⁻¹. Thus, global renewable water resources are estimated to range between 41 000 and 43 000 km³ yr⁻¹. The climate data sets WFD (available until 2001) and WFDEI (starting in 1979) were found to be inconsistent with respect to short wave radiation data, resulting in strongly different potential evapotranspiration. Global assessments of freshwater fluxes and storages would therefore benefit from the development of a global data set of consistent daily climate forcing from 1900 to current.

1 Introduction

Estimation of global scale freshwater fluxes, i.e. precipitation, runoff, river discharge, evapotranspiration and groundwater recharge is important for a variety of reasons. Precipitation is the only source of water for the continents. Knowledge of the amount and distribution of renewable water resources (long-term average runoff) for humans and environment is essential for a sustainable development (Hoekstra et al., 2012; Oki and Kanae, 2006). River discharge is of particular interest as it is a major source for water withdrawals, affects the habitat conditions of freshwater ecosystems and can be measured with high quality (accessible e.g. via GRDC database). Of interest are global amounts and spatial distribution of evapotranspiration (Jasechko et al., 2013; Jung et al., 2010; Sterling et al., 2012) as it affects energy and water transport in the atmosphere as well as precipitation. Estimation of groundwater recharge is equivalent to estimating renewable groundwater resources (Döll and Fiedler, 2008), and groundwater

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depletion (Wada et al., 2010). Assessments of climate change impacts on water resources and river discharge (Arnell and Gosling, 2013; Döll and Müller Schmied, 2012) or groundwater resources (Portmann et al., 2013) require reliable modeling of current and historic conditions.

5 There are different strategies to estimate global scale freshwater fluxes and storages. On the one hand, in-situ measurements can be interpolated if a dense monitoring network is available. Global precipitation products (e.g. GPCC (Schneider et al., 2014), CRU (Harris et al., 2013) and many more) were developed that way. Less dense point
10 measurements are used in combination with other sources for global assessments, e.g. evapotranspiration measurements from a network of flux towers together with remote sensing information (Jung et al., 2010). For estimating evapotranspiration, remote sensing data are used for deriving spatio-temporal input data for evapotranspiration equations (Miralles et al., 2011; Vinukollu et al., 2011; Wang and Liang, 2008). Total continental water storage variations can be estimated using GRACE gravity data
15 (Schmidt et al., 2006). Alternatively, land surface models (LSMs) and global hydrological models (GHMs) can be used to estimate spatio-temporal patterns of water fluxes and storages. GHMs are explicitly designed to assess the state of freshwater resources and to address water-related problems like floods and droughts (Corzo Perez et al., 2011; Prudhomme et al., 2011) and human impacts on freshwater resources. In the
20 last 20 yr, a number of GHMs have been developed using different conceptual approaches, e.g. VIC (Nijssen et al., 2001), WBM (Vörösmarty et al., 1998), Mac-PDM (Gosling and Arnell, 2011), WaterGAP (Alcamo et al., 2003; Döll et al., 2003) and PCR-GLOBWB (Sperna Weiland et al., 2010). First GHM and LSM model intercomparison experiments (Gudmundsson et al., 2012a, b; Haddeland et al., 2011; Van Loon et al.,
25 2012; Schewe et al., 2013) show large differences in simulated states and fluxes due to the model algorithms and parameters used, even if the climate input is identical.

As models are inherently imperfect, uncertainties exist. Comparing to the regional scale, epistemic uncertainty due to a lack of knowledge and understanding is of particular importance at the global scale (e.g. see discussion in Beven and Cloke, 2012;

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Wood et al., 2011, 2012). Three sources of uncertainty are generally defined, resulting from: model parameters, spatially distributed input data (e.g. climate forcing, water use, land cover) and model structure (or modeling approach). Hydrological processes are often represented in models by using effective parameters to transform the systems (non-)knowledge into solvable equations. These parameters are generally not measurable and, hence, are a source of uncertainty that can influence model results to varying degrees. Model calibration, i.e. adjustment of uncertain model parameters such that model output becomes similar to observations, is an approach to deal with parameter uncertainty. Basin-scale hydrological models are routinely calibrated against observed river discharge (e.g. Beven, 2001), while WaterGAP is the only GHM that is calibrated against observed river discharge in a basin-specific manner (Döll et al., 2003; Hunger and Döll, 2008; Werth and Güntner, 2010).

Uncertainties of climate forcings are fundamental. For example, Guo et al. (2006) showed the large sensitivity of soil moisture simulations of 11 LSMs to different climate forcings (esp. to precipitation and radiation) and concluded that this uncertainty on land surface hydrology is as large as the variations among the LSMs. Biemans et al. (2009) analyzed seven global precipitation products for 294 river basins worldwide and found out an average uncertainty of 30 % per basin. They studied the (uncalibrated) dynamic global vegetation and hydrology model LPJmL with those precipitation forcings and quantified an average uncertainty in discharge of about 90 %. Even though climate forcings are of such importance, only few studies are available reflecting this uncertainty in a global hydrological model setup.

Uncertainties in terms of model structure are related to the design of the model, i.e. the (number of) processes considered and their representation by conceptual approaches. To consider this kind of uncertainty, Clark et al. (2008) developed a framework to diagnose different structures of hydrological models, while approaches for addressing uncertainties related to model structure were developed by Butts et al. (2004), Refsgaard et al. (2006), and Song et al. (2011). Model intercomparison efforts in which identical climate forcings are used to drive all investigated models (e.g. WATCH

WaterMIP, ISI-MIP) show the effects of different model structures. For example, values for global annual evapotranspiration between 60 000 and 85 000 km³yr⁻¹ were reported in the WATCH WaterMIP study (Haddeland et al., 2011). However, in such multi-model studies, many completely different models are participating, which makes it very difficult to identify the reasons for different model behavior. A sensitivity study using basically the same model but with a refined model structure can therefore be of benefit (e.g. Thompson et al., 2013).

The goal of this study is to analyze the impact of the uncertainty related to spatially distributed input data and of model structure and modeling approach on water fluxes and storages at the global scale, using the most recent version of the GHM WaterGAP 2.2. Parameter uncertainty is neglected. The study was motivated by newly available climate forcing and land cover input data on the one hand, and the significant modifications regarding model structure of the state-of-the-art GHM WaterGAP (Alcamo et al., 2003; Döll et al., 2003) during the last decade on the other.

In particular, we will answer the following research questions:

1. How sensitive are freshwater fluxes and water storages to spatially distributed input data (climate forcing, land cover)?
2. What are the benefits of WaterGAP model structure refinements implemented during the last decade?
3. How does the modeling approach (calibration procedure, consideration of human water use) affect freshwater fluxes and water storages?
4. Which type of uncertainty is dominant for specific fluxes and variations of total water storage?

After an initial description of WaterGAP 2.2 (for details see the Appendix), the experimental setup is explained (Sect. 2). In Sect. 3, the results are described; focusing on the effect of the different model variants on global freshwater fluxes and water storages

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aspect, including either alternative climate forcing (CLIMATE), land cover input (LANDCOVER) or model structure (STRUCTURE). Each model variant was independently calibrated. Variant NoCal is an uncalibrated simulation with the standard version of WaterGAP 2.2 to study the impact of the calibration approach. Variant NoUse reflects naturalized water flows and storages without the impact of human water use, and thus also renewable water resources.

In addition, for assessing the effect of uncertainties on renewable water resources, variants CLIMATE, LANDCOVER, STRUCTURE and NoCal are also run without considering any water abstractions. The modeled time span is from 1901 to 2009. In this paper, model results for 1971–2000 are evaluated.

2.2.1 Climate input

Climate forcing data for global scale hydrological models are a major source of uncertainty for two main reasons: (1) they are subject to measurement errors which were not corrected in the original input data and (2) they are subject to interpolation errors due to low spatial and temporal monitoring network density and/or because (temporal) data gaps have to be filled. To analyze the sensitivity of different climate forcing datasets on calibration and subsequently on freshwater fluxes, two climate forcings were used to force both the WGHM and the Global Irrigation Model GIM (Döll and Siebert, 2002).

In variant STANDARD, the daily WATCH Forcing Data methodology applied to ERA-40 data (WFD) (Weedon et al., 2011) for the years 1901 to 1978 (the years 1901 to 1957 are based on reordered reanalysis data) and the WATCH Forcing Data methodology applied to ERA-Interim data (WFDEI) for the years 1979 to 2009 was chosen. WFD and WFDEI monthly sums/means are bias-corrected with other data sources (temperature bias correction, shortwave radiation adjustment using cloud cover and adjustment of number of wet days to CRU TS2.1 for WFD and to CRU TS 3.1 for WFDEI as well as adjustment of monthly precipitation sum to GPCC v4 (WFD) and GPCC v5 (WFDEI) and snowfall undercatch corrected after Adam and Lettenmaier, 2003). To calculate net shortwave radiation, the incoming shortwave radiation is reflected by literature based

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land cover specific albedo values (see Table A2). Literature based emissivity values for all land cover classes (Wilber et al., 1999) and the Stefan-Boltzmann-equation are used to calculate outgoing longwave radiation. The difference to incoming longwave radiation represents net longwave radiation. Net radiation is the sum of both components.

In variant CLIMATE, the monthly dataset CRU TS 3.2 (Harris et al., 2013) was used but monthly precipitation totals were replaced by the latest GPCC v6 precipitation monitoring product (Schneider et al., 2014) because it includes more observation stations. Neither CRU nor GPCC precipitation is corrected for observational errors, e.g. wind undercatch. Thus, Döll and Fiedler (2008) included the catch ratios of Adam and Lettenmaier (2003) and used the empirical function of Legates (1987) to correct especially snow undercatch by dividing snow and liquid precipitation using a temperature based approach. The correction of precipitation measurement bias leads to an average increase of 8.7% compared to the original product. On 37.5% of the land area (except Greenland and Antarctica), the increase of precipitation is larger than 10%. Differences of mean values from both datasets (CRU/GPCC and WFD/WFDEI) occur due to the slightly different precipitation correction approach and the GPCC version used for scaling monthly sums. Monthly precipitation is equally distributed to the number of wet days provided by the CRU 3.2 dataset; the distribution of wet days within a month is modeled as a two-state, first-order Markov chain (Döll et al., 2003). Cloudiness fraction was used to calculate incoming short wave radiation as well as outgoing long wave radiation after Shuttleworth (1993), see also Döll et al. (2003).

2.2.2 Land cover input data

The distribution of land cover classes and associated attributes are affecting simulated fluxes in terms of radiation energy balance (albedo and emissivity), snow dynamics (degree-day factor D_F), available soil water capacity (rooting depth) and interception capacity (L) (for details see the Appendix A). To estimate the effect of different, homogeneous-source land cover data, two input maps were used (Fig. 2). Attributes

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and model parameters associated to land cover classes were derived from literature or previous model versions (Table A2) and left equal in both variants.

In variant STANDARD we used the gridded MODIS (Moderate-resolution Imaging Spectroradiometer) land cover product (MOD12Q1) for the year 2004. The product MOD12Q1 (1 km resolution, global coverage up to 60° N) was used with land cover type 1 (IGBP classification). After resampling to 0.5° spatial resolution, the dataset was reclassified to fit to the WaterGAP land cover classification system (Table A2). As water bodies (GLWD) and percentage of urban area (from previous model versions) are obtained by additional input files, the second land cover class was appointed in case of “water” or “urban and built-up” as primary land cover. For coastal grid cells which are not fully covered by MODIS and north of 60° N, GLCC+CORINE land cover information was used.

In variant LANDCOVER, a combination of the Global Land Cover Characteristics database GLCC (USGS, 2008) based on the years 1992/1993 and, for Europe, CORINE Land Cover based on the year 2000 (European Environment Agency, 2004) was used as land cover information, as also in a previous WaterGAP version (Haddeland et al., 2011). The idea was to use an IGBP based classification scheme and a remote sensing based land cover distribution instead of IMAGE (Alcamo et al., 1998) model outputs (as in previous model versions). Both input datasets have a resolution of 1 by 1 km and were aggregated to the 0.5° model resolution by assigning the majority land cover type.

2.2.3 Structural model changes

During the last 10 yr, the WaterGAP model was subject to several revisions and improvements in terms of hydrologic process representation, resulting in an overall more complex model structure. To assess the sensitivity of simulated freshwater fluxes to model complexity, one model variant with a simplified structure comparable to Döll et al. (2003) (variant STRUCTURE) was set up which was run with the same input

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data as all other model variants. Differences of variant STRUCTURE as compared to STANDARD are as follows:

- Flow velocity is globally set to 1 ms^{-1} and the meandering ratio is set to 1, instead of the variable flow velocity algorithm of Verzano et al. (2012) in STANDARD.
- Reservoirs are treated as global lakes, i.e. the reservoir operation algorithm of Döll et al. (2009) is not used, which should result in a more dynamic discharge downstream of reservoirs.
- Human water demand is entirely satisfied from surface water resources, i.e. there are no groundwater abstractions as introduced by Döll et al. (2012).
- Evaporation from lakes/wetlands is not adjusted by reduction factors (Hunger and Döll, 2008) resulting in evaporation at potential rate even at low storage.
- Snow accumulation and melt are modeled on 0.5° (instead of the 3 arc minute sub-grid, Schulze and Döll, 2004) which should lead to less snow dynamics.
- Finally, there is no distinction in groundwater recharge for arid and humid regions (in contrast to Döll and Fiedler (2008) all regions are treated like humid regions) resulting in higher groundwater recharge in arid regions.

3 Results

3.1 Global water balance

Table 2 lists global values for various components of the global water balance and changes in total water storages (calculated excluding Antarctica, Greenland and inland sinks) as estimated by the different model variants. Global values vary mainly due to calibration and selected climate forcing. Global precipitation P is about $1900 \text{ km}^3 \text{ yr}^{-1}$ (or 1.7%) higher when using the CLIMATE model variant which results in an equal

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increase of discharge compared to STANDARD. Except for NoCal, global actual evapotranspiration AET (calculated as sum of E_c , E_{sn} , E_s and E_w , see Appendix A) does not vary considerably among the variants. In general, discharge to oceans and inland sinks is lower by the amount of change in AET. Actual water consumption WC_a (row 4 in Table 2) varies due to the demand of surface water abstractions and groundwater abstractions (which differs in CLIMATE due to the forcing of GIM (Appendix C) and in STRUCTURE where water demand is entirely extracted from surface water resources) and due to the different water availability for abstractions. In all cases, a large share of the total water demand could be satisfied (between 90 % in STRUCTURE and 96 % in CLIMATE).

When water use is not considered (NoUse), more evaporation ($131 \text{ km}^3 \text{ yr}^{-1}$) is modeled as there is more water available in the storages. Please note that additional evapotranspiration of irrigated crops is not included in AET but quantified within WC_a (row 4 in Table 2). As expected, river discharge is higher (by $758 \text{ km}^3 \text{ yr}^{-1}$) in NoUse. Changes in total water storages ($143 \text{ km}^3 \text{ yr}^{-1}$ less storage decrease) are also visible, especially due to no groundwater withdrawals in this variant (Table 3). The sum of these differences between STANDARD and NoUse is $1032 \text{ km}^3 \text{ yr}^{-1}$ which equals to WC_a (row 4 for STANDARD in Table 2).

The calibration has a strong effect on freshwater fluxes. Global discharge to oceans and inland sinks Q in NoCal is about $6400 \text{ km}^3 \text{ yr}^{-1}$ (or 15.7 %) higher than in STANDARD, meaning that the main effect of calibration is lowering discharge. In many river basins, the calibration parameter γ is higher than the value 1 globally used in NoCal which reduces the share of effective precipitation actually contributing to runoff. Consequently, AET is lower by nearly the same amount.

When comparing CLIMATE to STANDARD, P and Q are both increased by around $1900 \text{ km}^3 \text{ yr}^{-1}$ whereas global AET sums are nearly equal. When partitioning the increased Q into calibrated and uncalibrated grid cells, most additional Q (1546 of overall $1906 \text{ km}^3 \text{ yr}^{-1}$) is generated in non-calibrated grid cells mostly because of an increased

3.2 Actual evapotranspiration

Mean annual actual evapotranspiration AET shows the highest values around the equator consistent with available energy, except for the Pacific Rim of South America (Fig. 3a).

5 Among the variants, the largest differences to STANDARD occur in case of the uncalibrated version NoCal (Fig. 3f). As the calibration approach also affects grid cells outside of the 1319 calibration basins due to the regionalization (Appendix B3), all grid cells are affected. In most regions, calibration leads to higher AET, but in the upstream Amazon, the Congo, Arctic river basins and some other basins, the opposite is true. The global sum of AET of NoCal is 9.2 % lower than estimated with STANDARD (Table 2). Notable differences in AET also occur when using an alternative climate input (Fig. 3b). AET increases in CLIMATE on 42.6 % of the land surface by more than 10 mm yr⁻¹ and decreases by more than 10 mm yr⁻¹ on 30.5 % of the land surface. It increases (decreases) by more than 100 mm yr⁻¹ on 5.4 % (5.6 %) of the land surface. When summed globally, only minor changes in AET occur in case of CLIMATE (increase of 0.06 % or 39 km³ yr⁻¹, Table 2). In contrast, AET differences of the STRUCTURE variant are higher for the global sum (increase of 0.6 % or 414 km³ yr⁻¹) but occur on an overall smaller area (increase by more than 10 mm yr⁻¹ on 11.9 % of the land surface, decrease on 14.2 %). The effect of STRUCTURE is visible in areas with surface water bodies and in snow-dominated areas. On the one hand, an increase in net radiation in snowy regions leads to a slight increase of AET but in small absolute numbers as total AET is comparatively low. On the other hand, effects due to the evaporation reduction factor for surface water bodies are visible. In all variants except STRUCTURE, evaporation is limited when the surface water body storage is reduced to mimic the shrinking of surface area. Hence, in regions with a high percentage/volume of surface water bodies, AET is increased.

25 AET differences between LANDCOVER and STANDARD (Fig. 3c) are caused by changes in net radiation in energy-limited areas (not shown) as well as changes in

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rooting depth. In general, minor differences occur (except in some basins, see explanation below). In some regions, an increasing net radiation results in an increasing AET, e.g. in parts of Angola. In water-limited areas (e.g. north eastern Brazil), insignificant changes of AET occur even if net radiation strongly increases. In northern Australia, AET increases even when net radiation is reduced. Here, large parts are defined in STANDARD as open shrubland (rooting depth of 0.5 m) and in LANDCOVER as savanna (rooting depth of 1.5 m). As soil storage capacity is a function of rooting depth, even with more energy available for evapotranspiration, only half of the soil water can be evapotranspired due to the limited rooting depth. Neglecting human water abstraction in variant NoUse would lead to an overestimation of AET in regions where water abstraction for irrigation leads to reduction of wetlands areas (Fig. 3e), and a global AET overestimation by less than 0.2 % (Table 2).

In WaterGAP 2.2, AET can become negative in some (mostly snow dominated) regions, where precipitation input is too low to reproduce observed discharge (grey colors in Fig. 3a). The total water balance of each large water body is calculated in the outflow cell, hence AET can become very large as the value in mm is calculated by dividing AET over the whole lake by grid cell area. Moreover, in calibration basins, AET is adjusted in such a way that it is consistent with precipitation and simulated discharge and affected by correction factors CFA and CFS (calibration details see Appendix B).

3.3 Renewable water resources

Renewable water resources RWR (mean annual runoff of the grid cell to the river without consideration of human water use) are dominantly influenced by the calibration (NoCal) and subsequently by input data and model structure (Fig. 4).

As RWR are approximately the difference between precipitation and AET, the difference maps (Fig. 4b–e) represent more or less the inverted difference maps of Fig. 3 of the previous section. Compared to STANDARD, largest differences occur in model variant NoCal. In contrast to AET, calibration leads in many cases to lower RWR. The global sum of RWR of NoCal is 15.8 % higher than with STANDARD (Table 2). The global

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sum of RWR from CLIMATE is 4.7% higher but with large spatial spread. RWR decreases in CLIMATE on 21.4% of land surface by more than 10 mm yr^{-1} and increase by more than 10 mm yr^{-1} on 29.9% of the land surface. RWR decreases (increases) by more than 100 mm yr^{-1} on 4.7% (9.0%) of the land surface. The differences in LANDCOVER mainly follow differences in net radiation (not shown). In snow-dominated regions, RWR are lower in STRUCTURE because snow cover dynamics are less intense than in STANDARD. In grid cells with (large) surface water bodies, RWR are lower in STRUCTURE (as AET is unlimited here even if storages are nearly empty).

3.4 River discharge

3.4.1 River discharge seasonality

River discharge is the integral result of runoff generation, water losses by evaporation from surface water bodies, positive or negative net abstractions from surface water bodies and groundwater, and routing processes. It is one of the most important diagnostic variables in water resources. In many regions, river discharges have been observed for decades, providing an important data source for model evaluation. A good representation of modeled seasonality in comparison to the observed one is therefore a criterion for model evaluation. We compared observed and modeled discharge seasonality at the outflow of 12 large river basins, covering different climatic zones and levels of anthropogenic influence (Fig. 5). Climate input and model structure influence modeled discharge seasonality more than land cover changes for the selected river basins (NoCal is not shown as the Y-axis would have a very large spread). In the Mekong, for example, only marginal differences occur due to land cover because discharge is dominated by the climatic conditions (monsoon region). The simulated discharge of STRUCTURE gives the best fit in the Lena basin, indicating that (in this particular case) the revised snow algorithm worsened the representation of snow cover dynamics. The effect of many reservoirs in the Rio Parana is visible as the reservoir algorithm smoothes the discharge seasonality which fits better to the observation. In the

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Amazon, the amplitude is well reproduced by all variants but the peaks are shifted two months (in case of STRUCTURE) or one month. Even though there is an improvement due to model development, the storage dynamics are not perfect. For the Mackenzie River, all model variants are nearby but differ from observations. Here, freezing and thawing of the river are not reproduced as none of the model variants represents these processes. Interestingly, the Lena river basin is also frozen during winter time but here, low flows are simulated quite well. A strong impact of the climate input is found at the Danube River, where CRU TS 3.2/GPCC v6 forcing leads to much higher discharge amplitudes in comparison to WFD/WFDEI input and the observations. In many cases, climate and model structure affect the discharge seasonality most while differences between the land cover inputs are marginal.

3.4.2 Monthly time series

Nash–Sutcliffe efficiencies E_{NS} (Eq. 1, Nash and Sutcliffe, 1970) were calculated for time series of monthly river discharges at 1319 gauging stations used for calibration.

$$E_{NS} = 1.0 - \frac{\sum_{i=1}^n (O_i - S_i)^2}{\sum_{i=1}^n (O_i - \bar{O})^2} \quad (1)$$

with O_i is observed discharge, S_i is simulated discharge and \bar{O} is mean observed discharge (all units in $[\text{km}^3 \text{ month}^{-1}]$).

By adjusting the mean annual river discharge as done in our calibration approach, E_{NS} of monthly discharge increases in all calibrated model variants as compared to the NoCal variant, as E_{NS} is sensitive to both mean and variances (Fig. 6). Among all calibrated variants, STANDARD and NoUse achieve the highest mean E_{NS} values, while variant STRUCTURE shows a distinctly lower model performance (Fig. 6). This is further confirmed by the E_{NS} distribution per Köppen–Geiger region (Table 4, column

have higher seasonal variations than with water use because large return flows during the dry (irrigation) season smooth natural groundwater storage variations.

In addition, seasonal TWS variations in STRUCTURE differ from STANDARD particularly along large rivers (Fig. 7d), mostly with a smaller range in STRUCTURE. There, the flow velocity (variable in STANDARD) is lower than the constant 1 ms^{-1} in STRUCTURE, resulting in increased river storage. In many cold areas, the simpler snow algorithm in STRUCTURE leads to increased TWS seasonality.

4 Discussion

4.1 Comparison of simulated freshwater fluxes to other estimates

The modeled actual evapotranspiration AET and discharge to the oceans and inland sinks for all model variants are within the range of published values except the NoCal variant, which has very low AET and high discharge values (Tables 2 and 5). Difficulties with such comparisons can occur if different time spans are used. In addition, different land area is used, e.g. Mu et al. (2011) is based on remote sensing data and neglects bare land surfaces (their area: $109.03 \times 10^6 \text{ km}^2$) whereas Mueller et al. (2013) covers $130.92 \times 10^6 \text{ km}^2$ (which is also a reason for a larger AET).

Global discharge estimates differ due to the applied approach but are also highly dependent on the precipitation dataset used. The lowest discharge sum is published in Mueller et al. (2013), where global discharge is calculated as the difference between a mean of current available (but not precipitation undercatch corrected) precipitation products and AET. Compared to other estimates (and Mueller et al., 2011), AET is low. In addition, the assumed mean precipitation of $\sim 99\,000 \text{ km}^3 \text{ yr}^{-1}$ is low compared to recent estimates of Schneider et al. (2014) which are about $117\,000 \text{ km}^3 \text{ yr}^{-1}$ or the values used in this study (Table 2). When comparing the presented model variants with previous estimates of WaterGAP, model refinements lead to an increase of discharge (STANDARD is approx. $450 \text{ km}^3 \text{ yr}^{-1}$ higher than STRUCTURE, Table 2),

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which is consistent to lower numbers e.g. in Döll and Fiedler (2008). Moreover, when precipitation is not corrected as in Döll et al. (2003), discharge is even lower.

4.2 Benefits and limitations of the calibration approach

The applied calibration approach is clearly beneficial as it leads to a better fit of simulated to observed monthly river discharge time series (Fig. 6 and Table 4). Consequently, the basin-specific adjustment of 1–3 parameters (γ , CFA and CFS, see Appendix B2) based on observed mean annual discharge has been part of the WaterGAP modeling approach since the beginning. Calibration allows to a certain degree compensating errors in input data and effective model parameters. Also, structural problems of the model, e.g. due to the simplified representation of hydrological processes at a half-degree grid cell, may be balanced out. The effect of calibration on modeled renewable water resources (Fig. 4e) dominates all other modifications within this study setup.

However, the correction of total cell runoff using CFA and CFS that is required to force simulated mean annual river discharge values to be equal to observed values is not ideal and has undesirable effects on estimated AET and RWR. For example, at the river basin Yenisey at station Igarka (Western Siberian Plain), one half of the basin has strongly reduced and the other half strongly increased AET when using alternative climate forcing. Transferring the correction factor CFS (which is, if necessary, calculated at the outflow grid cell of the basin) to the upstream grid cells can lead to unrealistic high positive and negative values for AET if precipitation is too low in these parts of the basin to simulate observed discharge or the AET of surface water bodies has to be reduced by CFA. This is the reason for some artificial patterns in Fig. 3 and consequently in Fig. 4. These kinds of consistency errors can be found in some more basins where cumulative AET is low and parts of the basins are covered with surface water bodies. Nevertheless, the approach ensures a closed water balance for the whole basin.

Obviously, one parameter is not sufficient to calibrate the model. In many basins the γ parameter is not sensitive to input data and model structure in the current calibration approach as the range of γ through all four variants (NoCal is not considered, NoUse

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has the same value as STANDARD) is rather small. 59% of the basins in Fig. 8 are colored dark blue which means that the calibration parameter γ has the same value in all model variants. Here, γ is at its artificial boundaries minimum (0.1) or maximum (5.0) value and the influence of input data and model structure, which were modified in this study, is insignificant. On the other hand, in 21 % of the basins, γ is differing by > 1 (green, yellow and red colors). In these basins, the calibration parameter is sensitive to input data and model structure. Anyhow, within future model development, one task is to restructure the calibration approach with the aim to avoid correction factors or rather to introduce and test alternative calibration objectives. This could be achieved by either including more parameters (multi-parameter calibration) and/or by integrating additional reference data, e.g. GRACE based data as was shown by Werth and Güntner (2010) (multi-objective calibration). In addition, remote sensing based input data with global coverage have been available for a decade. Especially for land cover characteristics (e.g. land cover type, L , albedo, see Appendix A), a more realistic representation of dynamics (integration of time series as input data instead of static input maps) can reduce the input data and model parameter uncertainty.

4.3 How sensitive are freshwater fluxes and water storages to spatially distributed input data (climate forcing, land cover)?

In general, more differences occur due to the alternative climate input than due to the alternative land cover data. The major freshwater fluxes (AET, Fig. 3 and RWR, Fig. 4) as well as river discharge (Fig. 5) show in many cases that land cover input has much less impact (except for some areas where the attributes of a changed land cover type differ significantly). Forced with CRU 3.2 and GPCC v6 instead of WFD/WFDEI input, AET is increased by at least 10 mm yr^{-1} in large parts in the world (light blue colors in Fig. 3b). In some regions, RWR decreases by the same amount (e.g. South East Asia, Australia, Saudi Arabia), while in others no clear effect on RWR is detectable (e.g. North America). In some parts of Europe, RWR increases by at least 10 mm yr^{-1} even

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(Hunger and Döll 2008) and the introduction of a bias-correction for observed precipitation (Döll and Fiedler, 2008) has had the problematic consequence that correction factors to lower simulated river discharge have increasingly been required to ensure that simulated mean annual river discharges are equal to observed values.

5 4.5 How does the modeling approach (calibration procedure, consideration of human water use) affect freshwater fluxes and water storages?

The calibration procedure reduces simulated river discharge and water resources on most of the land area and increases the AET (Figs. 3 and 4, Table 2). Without calibration, global AET and discharge would rank at the lower and higher end of the published values, respectively (Table 5). In addition, the fit to observed monthly river discharge time series as quantified using the E_{NS} criterion would worsen almost everywhere (Fig. 9f). The impact of calibration on freshwater fluxes and water storages is higher than those of alternative climate forcings and land cover data, and of a more sophisticated model structure. This confirms the strong benefit of calibration. However, as E_{NS} is affected by mean discharge as well as discharge variations, the calibration approach improves this criterion.

Compared to the other variants, the consideration of human water use does not have large effects on freshwater fluxes and storages at the global scale. However, in regions with intense water use, in particular from surface water bodies (e.g. in Pakistan), AET is reduced by human water use (Fig. 3e). When surface water storage decreases due to water use, the reduction factor (Appendix A4) decreases evaporation from surface water bodies. Please note that additional evapotranspiration of irrigated crops is not included in AET but quantified as WC_a (Table 2). If the impact of human water use on river discharge were not considered, van Beek et al. (2011) showed lower performance in general. Within our study, higher correction factors would be necessary in basins with large abstractions from surface water bodies or significant decreases of baseflow due to groundwater abstractions. Still, E_{NS} of basins with high amounts of human water use is generally lower than those without human water use (not shown).

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In some basins mainly in northeastern Europe, E_{NS} improves when neglecting human water use (Fig. 9e). This obviously reflects uncertainties in water use models.

4.6 Which type of uncertainty is dominant for specific fluxes and variations of total water storage?

The answer to this question depends on the type of fluxes and the spatial aggregation. Regarding selected global sums of freshwater fluxes (discharge Q into oceans and inland sinks and actual evapotranspiration AET) and mean annual total water storage trends $dTWS$, dominant uncertainties can be determined by computing differences between the values computed with certain model variant and STANDARD. As already shown above, global values of AET and Q as well as the fit of simulated to observed river discharge time series (E_{NS}) are most sensitive to whether the model is calibrated or not (Table 6). STRUCTURE and NoUse have the strongest impact on the global TWS trend (Table 6) as these model variants cannot reflect groundwater depletion. More refined model algorithms rank second regarding global AET sums and E_{NS} , and alternative climate forcings rank second regarding river discharge and third regarding median E_{NS} . The alternative land cover input data sets have the overall lowest impact on computed freshwater fluxes and storages.

Regarding grid cell-specific differences that are more relevant than global values for most applications, the ranking of dominant uncertainties is quite different. Patterns of seasonal TWS variations are affected most strongly by the climate forcing (Fig. 7b), while climate forcings show the second largest impact on the spatial distribution of AET and RWR, after calibration (Figs. 3 and 4). The fraction of the global land area that is affected by significant differences of AET and river discharge between a certain model variant and the STANDARD variant is largest in case of NoCal, followed by CLIMATE, LANDCOVER, STRUCTURE and NoUse. Thus, both global and grid cell values are most sensitive to calibration. The larger sensitivities to climate forcings and land cover input at the grid cell level (Table 7) cancel when globally averaged. The larger sensitivities of globally aggregated values (Table 6) to structural changes and

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a combination of the monthly CRU 3.2 and GPCC v6 data sets as done for model variant CLIMATE. However, we found that it is problematic to combine the WFD climate data set (covering 1901–2001) with the only seemingly consistent WFDEI data set (covering 1979–2009) due to a radiation bias (short wave downward radiation component) between the two data sets. This results in a steep increase of potential evapotranspiration in 1979, and a water storage decrease between 1971 and 2000 that is an artifact of the combination of the two climate data sets (comp. Sect. 3.1). It would be very beneficial for an improved estimation of global freshwater fluxes and storages to have a consistent daily climate forcing that covers the whole 20th and the early 21st century.

The calibration approach of WaterGAP is necessary to compensate uncertainties of spatially distributed input data, parameters and model structure. However, a calibration of only one parameter related to soil water balance is not sufficient and correction factors have to be applied in a number of basins. Therefore, a redesign of the calibration approach, with additional observations (e.g. including TWS variations as derived from GRACE gravity fields), other calibration objectives and adjustment of more model parameters (without correction factors) is planned.

The improved representation of hydrological processes of WaterGAP within the last decade led to a more complex model structure but also to a better fit to observed river discharge in most cases. However, in some parts of the world, model performance is still not satisfactory due to an inappropriate modeling of certain processes such that further changes of the model structure are required. For example, the modeled discharge seasonality in the Amazon basin is shifted compared to the observed one, which is suspected to be caused by inappropriate modeling of the temporal variations of inundations and the neglect of backwater effects. The reservoir operation algorithm does not yet take into account the construction year of the dam. Moreover, model results in semi-arid and arid regions are weak, and improved modeling of evaporation from ephemeral ponds is planned.

Appendix A describes the WaterGAP Global Hydrology Model (WGHM) in its current version 2.2. In the order of processing, the single storage compartments and belonging in- and outflows are explained. Appendix B provides information on the calibration and regionalization approach WaterGAP is based on. Appendix C gives a brief introduction of the water use sub-models, and the GWSWUSE module is described in Appendix D.

Appendix A

Description of the WaterGAP Global Hydrology Model (WGHM)

A1 Canopy

The change of canopy storage S_c [mm] over time t [d^{-1}] is calculated as

$$\frac{dS_c}{dt} = P - P_t - E_c \quad (A1)$$

where precipitation P [$mm d^{-1}$] is the inflow and the amount of throughfall P_t [$mm d^{-1}$] and canopy evaporation E_c [$mm d^{-1}$] are the outflows.

Throughfall P_t is calculated as

$$P_t = \begin{cases} P & S_c \geq S_{c,max} \\ 0 & S_c < S_{c,max} \end{cases} \quad (A2)$$

Following Deardorff (1978), canopy evaporation E_c is calculated as

$$E_c = E_p \left(\frac{S_c}{S_{c,max}} \right)^{\frac{2}{3}} \quad (A3)$$

where E_p is potential evapotranspiration [$mm d^{-1}$].

E_p is calculated according to the Priestley–Taylor model (Priestley and Taylor, 1972), differentiating atmospheric water demand between humid ($\alpha = 1.26$) and semi-arid/arid ($\alpha = 1.74$) areas. Grid cells were defined as semi-arid/arid if long term average (1971–2000) precipitation is less than $0.5 \times E_p$ (UNEP, 1992).

5 S_c is limited between 0 and maximum canopy storage $S_{c,max}$, which is calculated as

$$S_{c,max} = m_c L \quad (A4)$$

where m_c is 0.3 [mm] and L is the leaf area index [-]. L is calculated based on a modified growth model described in Kaspar (2003) and is limited to minimum and maximum values. Maximum L values per land cover class (Table A1) are based on literature
 10 (Schulze et al., 1994; Scurlock et al., 2001). Minimum L values per land cover class are calculated as:

$$L_{min} = 0.1 f_{d,lc} + (1 - f_{d,lc}) c_{e,lc} L_{max} \quad (A5)$$

where $f_{d,lc}$ is the fraction of deciduous plants [-] and $c_{e,lc}$ is the reduction factor for evergreen plants [-] (Table A1). Development of L is simulated as a function of daily
 15 temperature and precipitation. The growing season starts when the daily temperature is above 8 °C for a land cover specific number of days (Table A1) and cumulative precipitation is at least 40 mm. During the growing season, L increases linearly until it reaches L_{max} after 30 days. In semi-arid and arid regions, it is necessary that at least 0.5 mm daily precipitation occurs to keep the growing season ongoing. If the condition
 20 for growing season is not fulfilled anymore, the senescence phase is initiated, i.e. L is degraded to L_{min} linear within 30 days.

A2 Snow

The change of snow water storage S_{sn} [mm] over time t [d⁻¹] is calculated as

$$\frac{dS_{sn}}{dt} = P_{sn} - M - E_{sn} \quad (A6)$$

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where P_{sn} is precipitation, falling as snow at temperatures below 0°C [mm d^{-1}], M is snow melt [mm d^{-1}] and E_{sn} is sublimation [mm d^{-1}].

Snow accumulation and melt are modeled on a 3 arc minute sub-grid (100 sub-grid cells per 0.5°) using a degree day algorithm (Schulze and Döll, 2004). Mean sub-grid elevation was derived from GTOPO30 (US Geological Survey, 2003). The daily temperature for each sub-grid cell is calculated from the temperature of the 0.5° cell, applying an adiabatic lapse rate of 0.6°C per 100 m. To avoid excessive snow accumulation, temperature does not decrease if a snow water equivalent of 1000 mm is reached in one sub-grid.

At temperatures below 0°C , all precipitation is assumed to fall and accumulate as snow. At sub-grid temperatures T [$^{\circ}\text{C}$] above melting temperature T_m (0°C) and if snow storage is present, snow melts with land cover specific degree-day factor D_F [$\text{mm d}^{-1}\text{C}^{-1}$] (Table A2) as:

$$M = \begin{cases} D_F(T - T_m) & T > T_m, \quad S_{sn} > 0 \\ 0 & \text{other} \end{cases} \quad (\text{A7})$$

Instead of using one specific albedo for snow as in previous versions ($\alpha = 0.4$), land cover specific snow albedo values are used to account for differences in reflective properties between the land use classes under snow-covered conditions (Table A2). The albedo value switches to snow albedo if snow water equivalent of the grid cell exceeds 3 mm, i.e. a closed snow cover is assumed. Sublimation E_{sn} is modeled like potential evaporation rate but applying a latent heat of 2.835 [MJ kg^{-1}] for temperatures below 0°C and $2.501 - 0.002361 \times T$ [MJ kg^{-1}] above 0°C .

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A3 Soil

Like snow and canopy, the change of soil water storage S_s [mm] over time t [d^{-1}] is calculated as one layer as:

$$\frac{dS_s}{dt} = P_{\text{eff}} - R_l - E_s \quad (\text{A8})$$

- 5 with effective precipitation P_{eff} [mm d^{-1}] as inflow and runoff from land R_l [mm d^{-1}] and actual evapotranspiration E_s [mm d^{-1}] as outflows.

$$P_{\text{eff}} = P_t - P_{\text{sn}} + M \quad (\text{A9})$$

with P_t is through fall [mm d^{-1}], (see Fig. A1), P_{sn} is precipitation falling as snow [mm d^{-1}] and M is snow melt [mm d^{-1}].

- 10 Actual evapotranspiration from the soil E_s [mm d^{-1}] is a function of potential evapotranspiration from the soil E_p [mm d^{-1}] minus the already evaporated water from the canopy E_c [mm d^{-1}], actual soil water content in the effective root zone S_s [mm] and total available soil water capacity $S_{s,\text{max}}$ [mm] as

$$E_s = \min \left((E_p - E_c), (E_{p,\text{max}} - E_c) \frac{S_s}{S_{s,\text{max}}} \right) \quad (\text{A10})$$

- 15 where $E_{p,\text{max}}$ is 20 mm d^{-1} in semi-arid and arid regions whereas 10 mm d^{-1} in grid cells classified as humid, $S_{s,\text{max}}$ is the product of total available water capacity in the upper meter of the soil (Batjes, 1996) and the land cover specific rooting depth (Table A2).

Runoff from land R_l [mm d^{-1}] is calculated after Bergström (1995) as

$$R_l = P_{\text{eff}} \left(\frac{S_s}{S_{s,\text{max}}} \right)^y \quad (\text{A11})$$

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Dependent on the soil water storage S_s , a part of effective precipitation P_{eff} becomes runoff. If the soil water storage is empty, $R_1 = 0$. If the soil is completely saturated (at $S_{s,\text{max}}$), runoff equals effective precipitation. Between these points, the runoff coefficient γ determines the amount of precipitation that converts to runoff. This parameter is used for calibration (see Appendix B1). In urban areas (defined as separate input map from IMAGE 2.2), 50 % of P_{eff} is directly passed to the river.

A4 Groundwater

Inflow to groundwater storage S_g [mm] is groundwater recharge R_g [mm d^{-1}], whereas outflows are baseflow Q_g [mm d^{-1}] and net abstractions from groundwater NA_g [mm d^{-1}] (Appendix C), which can also act as inflow (e.g. as additional groundwater recharge due to irrigation with surface water).

$$\frac{dS_g}{dt} = R_g - Q_g - NA_g \quad (\text{A12})$$

Groundwater recharge R_g [mm d^{-1}] is calculated as a fraction of runoff from land:

$$R_g = \min(R_{g,\text{max}}, f_g R_1) \quad (\text{A13})$$

where $R_{g,\text{max}}$ is soil texture specific maximum groundwater recharge [mm d^{-1}] (with values of 7/4.5/2.5 for sandy/loamy/clayey soils) and f_g is the groundwater recharge factor (ranging between 0 and 1) related to relief, soil texture, aquifer type and the existence of permafrost or glaciers. For a detailed description see Döll and Fiedler (2008). If a grid cell is defined as arid and has coarse (sandy) soil, groundwater recharge will only occur if precipitation exceeds a critical value of 12.5 mm d^{-1} . Both values, $R_{g,\text{max}}$ and the precipitation threshold, are adapted to the climate forcing used (WFD) aiming to reach comparable groundwater recharge patterns of (Döll and Fiedler, 2008) as that groundwater recharge estimation is confirmed with experts within the WHYMAP

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(<http://www.whymap.org>) efforts. Within CLIMATE, the original values 5/3/1.5 for $R_{g,max}$ and 10 mm d^{-1} as precipitation threshold were used.

The outflow is modeled with $k_g = 0.01 \text{ d}^{-1}$ as

$$Q_g = k_g S_g \quad (\text{A14})$$

- 5 The runoff from land R_l which is not groundwater recharge R_g , represents the fast surface runoff R_s and is routed, together with Q_g , through a series of different storages representing wetlands, lakes and reservoirs until reaching the river segment (Fig. A1).

A5 Surface water bodies

10 Surface water bodies (inland freshwater such as wetlands, lakes and reservoirs) play an important role in the hydrologic cycle e.g. for evaporation and the lateral transport. In general, surface water body storages S [m^3] increase by inflow I [$\text{m}^3 \text{ d}^{-1}$] from other storages or from upstream (see Fig. A1), and are reduced by the outflow Q [$\text{m}^3 \text{ d}^{-1}$]. Additionally, the water balance of the water body itself B [$\text{m}^3 \text{ d}^{-1}$] is calculated as $B = P - E_w$, where P is precipitation [$\text{m}^3 \text{ d}^{-1}$] and E_w is potential evaporation of open water
15 surfaces [$\text{m}^3 \text{ d}^{-1}$] applying an albedo of 0.08. Finally, net abstractions of surface water NA_s [$\text{m}^3 \text{ d}^{-1}$] are considered, resulting in the storage equation:

$$\frac{dS}{dt} = I - Q + B - NA_s \quad (\text{A15})$$

20 Outflow is in principle modeled like groundwater outflow (Appendix A4) for “local” lakes and wetlands, whereas “global” lakes and wetlands are linear storages whose equations are solved analytically.

WaterGAP 2.2 does not consider variable land/water fractions as would be expected when a lake is shrinking due to evaporation and land surface increases; thus Hunger and Döll (2008) introduced a reduction parameter which reduces the evaporation when lake/wetland storage is low. In WaterGAP 2.2, for all surface water bodies the reduction

factor r [-] is calculated as

$$r = 1 - \left(\frac{|S - S_{\max}|}{S_{\max}} \right)^p \quad (\text{A16})$$

where S is actual water body storage [m^3], S_{\max} is maximum water body storage [m^3] and p is the reduction exponent [-]. As no truly global dataset on lake volumes is available, the maximum storage capacity is determined by multiplying the surface area with an “active” depth (set to 5 m and 2 m for lakes and wetlands, respectively). Values for p are 3.32 for lakes and wetlands which means a reduction of evaporation by 10 % if storage is halved and 2.81 for reservoirs, which means a reduction of 15 % if storage is half of the maximum storage capacity (and a reduction of 50 % if storage is reduced to 20 % of storage capacity).

The distribution of wetlands is derived from GLWD (Lehner and Döll, 2004) as percentage of cell coverage. Locations and attributes of lakes and reservoirs are based on a combination of GLWD and a preliminary version of the GRanD database (Döll et al., 2009; Lehner et al., 2011). In total, 6553 reservoirs, 52 regulated lakes (lakes whose outflow is regulated by a dam) (from GRanD) and 242 798 unregulated lakes (from GLWD) were considered. Out of these, 1386 large lakes (area $\geq 100 \text{ km}^2$), 1110 large reservoirs (storage capacity $\geq 0.5 \text{ km}^2$) and 52 regulated lakes (area $\geq 100 \text{ km}^2$ or storage capacity $\geq 0.5 \text{ km}^2$) were classified as “global”, i.e. they receive inflow not only from the grid cell itself but also from upstream (“global” wetlands are defined in the same way, see Fig. A1). All other surface water bodies were classified as “local”. If “global” lakes or reservoirs cover more than one grid cell, the water balance of the whole surface water body is calculated at the outflow cell.

A6 Lateral routing

The global drainage direction map DDM30 (Döll and Lehner, 2002) is used to route the discharge through the stream network until it reaches the ocean or an inland sink.

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Fast runoff $R_s = R_l - R_g$ is routed to the surface storages without any delay, whereas baseflow Q_g is a function of groundwater storage (Fig. A1, Appendix A4). Due to limited information on groundwater flow between grid cells, the groundwater recharge can only contribute to groundwater runoff of the same grid cell.

Verzano et al. (2012) improved the routing by introducing a variable flow velocity approach based on the Manning–Strickler equation. The roughness coefficient is calculated after Cowan (1956) by using different physiographic parameters and information about rural and urban areas. The hydraulic radius is calculated using actual discharge of the cell and empirical relationships of river width and depth at bankfull flow conditions. Bankfull conditions are assumed to correspond to the 1.5 yr maximum series annual flow (Schneider et al., 2011) and were accordingly calculated from daily discharge time series for the global land surface. River bed slopes were calculated based on the HydroSHEDS drainage direction map (Lehner et al., 2008) and a meandering ratio (method is described in Verzano et al., 2012).

The reservoir algorithm of Hanasaki et al. (2006), distinguishing irrigation and non-irrigation reservoirs and considering 1109 reservoirs was implemented and improved by Döll et al. (2009) and slightly adapted in WaterGAP 2.2: if reservoir storage falls below 10 % of storage capacity, the release coefficient is set to 0.1 instead of 0.0 in Döll et al. (2009), assuring that at least some water is released e.g. for downstream ecosystem demands.

Appendix B

Calibration and regionalization

B1 Calibration approach

WGHM is calibrated against mean annual discharge by adjusting the runoff coefficient γ (Eq. A11) for all grid cells of each calibration basin and – if necessary – two

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additional correction factors. The calibration procedure of WGHM is well described in Döll et al. (2003) and Hunger and Döll (2008). As WaterGAP was developed to quantify water resources and water stress, calibration forces simulated discharge to be, during the calibration period, between 99 and 101 % of observed river discharge. It is implicitly assumed that the model should be robust enough to reproduce intra- and interannual variability. Main reasons for calibration are the uncertainty of input data, parameters and model structure as well as the scale of the model and grid cell heterogeneity. To overcome overparameterization and to keep the calibration as simple as possible, calibration is performed by adjusting the one free parameter γ (Eq. A11) within the limits 0.1 and 5.0. With low γ , runoff is high even if the soil is at low saturation, and with a high value, runoff is small even with nearly saturated soils. However, in many basins, adjustment of the soil water balance alone does not lead to a fit of simulated discharge to observed discharge for various reasons. These include uncertainty of climate forcing, underestimation of evaporation losses in dry areas caused by neglecting formation of ephemeral ponds and neglecting of streambed losses. In these cases, the area correction factor (CFA) is computed, which adjusts net cell runoff of each cell in the sub-basins. With limits between 0.5 and 1.5, cells with positive (precipitation > evapotranspiration) and negative (water body evapotranspiration > precipitation, e.g. global lakes which are fed by upstream inflow) are multiplied with a value symmetric around 1.0 (Hunger and Döll, 2008). In some basins, however, the adaptation of both γ and CFA is not sufficient for a successful calibration, i.e. the deviation between simulated and observed long term average discharge remains larger than 1 %. Possible reasons are discussed in Hunger and Döll (2008). To avoid error propagation to next the downstream basin, the modeled discharge is corrected to the measured discharge in the grid cell where the discharge station is located by multiplying with the station correction factor CFS (Hunger and Döll, 2008).

B2 Discharge stations used

Observed discharge time series were provided by the Global Runoff Data Center (GRDC). Following Hunger and Döll (2008), gauging stations listed in the GRDC catalogue (<http://grdc.bafg.de/>, download date: 28.09.2012) were included in the calibration setup if they fulfilled three main criteria: (1) an upstream area of at least 9000 km², (2) a time series of at least four (complete) years, and (3) an inter-station catchment area of at least 30 000 km². All in all, a number of 1319 stations, covering 53.6% of the global land area except Antarctica and Greenland, was used for calibration (Fig. B1). If available, the 30 yr period 1971 to 2000 was chosen as calibration years.

B3 Regionalization

In order to transfer the calibrated γ values to ungauged basins, the parameter is regionalized using a multiple linear regression approach relating the natural logarithm of the calibrated γ values to the following basin descriptors: mean annual temperature, mean available soil water capacity, fraction of open water bodies, mean basin land surface slope, fraction of permanent snow and ice, and the aquifer-related groundwater recharge factor. Like in calibration basins, the regionalized parameter values are constrained to the range 0.1 to 5.0. CFA and CFS are not regionalized but are set to 1.0 in uncalibrated basins.

Appendix C

Description of water use models

In pre-processing steps to the WGHM, the Global Water Use sub-models (left side of Fig. 1) provide water withdrawal and water consumption (the part of withdrawn water that is not returned to the system but evaporated or incorporated in products) for five

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sectors: irrigation, livestock farming, domestic use (households and small businesses), manufacturing industries and thermal power plant cooling.

Irrigation water consumption is calculated on daily time steps for each grid cell by the Global Irrigation Model (GIM) on the basis of gridded area equipped for irrigation (Siebert et al., 2005, 2007) and climate as full irrigation (the difference between potential evapotranspiration and effective precipitation) of paddy rice and non-rice crops, based on modelled cropping patterns (Döll and Siebert, 2002). Water withdrawals for irrigation are calculated by dividing consumptive use by country-specific irrigation water use efficiencies (Rohwer et al., 2007). Consumptive livestock water use is calculated as a function of animal numbers per grid cell and water requirements per capita for ten different livestock types, while national values of domestic and manufacturing water use are downscaled to the grid cells using population density (Flörke et al., 2013). Cooling water use per grid cell accounts for the location of more than 60 000 power plants, their cooling and combustion type, and their electricity production (Flörke et al., 2013; Vassolo and Döll, 2005). Temporal development of domestic, manufacturing, and cooling water use is calculated as water use intensity per capita or unit industrial output (considering structural and technological change over time), multiplied by the driving force of water use, either population (for domestic use), national manufacturing output (as Gross Value Added, which is a share of Gross Domestic Product), or national thermal electricity production (Flörke et al., 2013). While WGHM uses aggregated monthly time series of irrigation consumptive use, the other sectoral water uses are distributed equally throughout the year.

Appendix D

Description of GWSWUSE

Döll et al. (2012) introduced sector-specific groundwater use fractions for irrigation, domestic and manufacturing water use, based on census information and data collected

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by the International Groundwater Resources Assessment Centre (IGRAC) (Siebert et al., 2010). In WaterGAP, water abstraction for livestock water use and cooling of thermal power plants are assumed to be from surface water only. Based on information from the five water use models, GWSWUSE first calculates consumptive irrigation water use for time series of irrigated area per country, and the respective irrigation water withdrawals from surface water and groundwater. Finally, GWSWUSE computes, for each grid cell, net abstraction from surface water and net abstractions from groundwater, taking into account return flows. Due to return flows, net abstractions can be positive (water is abstracted from storage) or negative (water is added to storage).

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Table 2. Long-term average (1971–2000) freshwater fluxes from global land area (except Antarctica and Greenland) of WaterGAP 2.2 in $\text{km}^3 \text{yr}^{-1}$. Cells representing inland sinks were excluded but discharge into inland sinks was included.

nr	component	STANDARD	NoUse ^f	CLIMATE	LANDCOVER	STRUCTURE	NoCal
1	precipitation P	111 070	111 070	112 969	111 070	111 070	111 070
2	actual evapotranspiration AET	69 803	69 934	69 842	70 012	70 217	63 344
3	discharge into oceans and inland sinks Q^a	40 458	41 216	42 364	40 250	40 002	46 822
4	water consumption (actual) (rows 5 + 7) WC_a	1031	0	927	1029	983	1054
5	net abstraction from surface water (actual) ^b	1102	0	960	1102	983	1126
6	net abstraction from surface water (demand) NA_s^c	1154	0	1000	1154	1082	1154
7	net abstraction from groundwater NA_g^d	-72	0	-33	-72	0	-72
8	change of total water storage dS/dt^e	-215	-73	-156	-214	-44	-143
9	long term averaged yearly volume balance error ($P - AET - Q - WC_a - dS/dt$) deviation to P	-7	-7	-8	-7	-88	-7
		-0.006 %	-0.006 %	-0.007 %	-0.006 %	-0.08 %	-0.006 %

^a including anthropogenic water use (except NoUse),^b if not enough water is available, demand is not completely satisfied,^c demand that needs to be satisfied (water use model output),^d negative due to return flows when irrigating with surface water; groundwater demand is always satisfied,^e total water storage (TWS) of 31. December 2000 minus TWS of 31. December 1970 divided by 30 yr,^f STANDARD but no subtraction of water use; discharge into oceans and inland sinks equals renewable water resources.

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Table 3. Changes in freshwater storage compartments (except Antarctica and Greenland) between 31. December 1970 and 31. December 2000 in $\text{km}^3 \text{yr}^{-1}$. Cells representing inland sinks were excluded.

compartment	STANDARD	NoUse ^a	CLIMATE	LANDCOVER	STRUCTURE	NoCal
total water storage	-214.8	-73.7	-156.4	-214.8	-44.5	-143.0
canopy	-0.05	-0.05	0.002	-0.05	-0.05	-0.05
snow	-3.0	-3.0	-6.3	-3.3	-1.3	-3.0
soil	-21.6	-21.6	-0.9	-20.6	-20.9	-20.0
groundwater	-124.9	8.6	-126.9	-125.4	9.7	-82.7
local lake	-1.9	-1.5	-0.3	-1.9	-2.1	-1.1
local wetland	-4.9	-4.3	1.9	-5.1	-8.4	-2.2
global lake	-3.5	-3.4	-1.1	-3.4	-8.2	-3.8
reservoirs	-43.1	-37.5	-23.2	-43.1	*	-21.4
global wetlands	-4.9	-4.3	1.9	-5.1	-8.4	-2.2
river	-6.7	-6.0	2.7	-6.7	-4.6	-4.3

^a STANDARD but no subtraction of water use.

* not applicable as reservoirs are treated as global lakes.

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**Table 4.** Number of calibration basins per E_{NS} category and Köppen–Geiger climate zone^a.

Variant	class	E_{NS}	A	B	C	D	E	sum
STANDARD	1	> 0.7	75	19	117	129	29	369
	2	0.5–0.7	100	17	68	134	18	337
	3	< 0.5	110	91	83	282	47	613
CLIMATE	1	> 0.7	67	8	77	145	30	327
	2	0.5–0.7	116	31	68	107	26	348
	3	< 0.5	104	79	127	293	41	644
LANDCOVER	1	> 0.7	77	20	117	128	32	374
	2	0.5–0.7	94	16	68	132	15	325
	3	< 0.5	114	91	83	285	47	620
STRUCTURE	1	> 0.7	63	20	85	99	27	294
	2	0.5–0.7	101	16	84	132	22	355
	3	< 0.5	121	91	99	314	45	670
NoUse	1	> 0.7	77	15	109	138	30	369
	2	0.5–0.7	97	26	68	130	17	338
	3	< 0.5	111	86	91	277	47	612
NoCal	1	> 0.7	17	5	39	61	12	134
	2	0.5–0.7	28	4	32	80	11	155
	3	< 0.5	240	118	197	404	71	1030

^a Calculated by WaterGAP after (Kottek et al., 2006); A: equatorial climate, B: arid climate, C: warm temperate climate, D: snow climate and E: polar climates. Note that the number of basins per climate zone differs for CLIMATE as here, the basis for Köppen–Geiger climate calculation is CRU TS 3.2 and GPCC v6 instead of WFD/WFDEI climate input for all other variants.

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Table 5. Comparison of diverse estimates of global actual evapotranspiration and discharge in $\text{km}^3 \text{yr}^{-1}$.

actual evapotranspiration		discharge	
62 800	Mu et al. (2011)	34 406	Mueller et al. (2013)
64 512 ^a	Mueller et al. (2013)	36 200	Wada et al. (2010)
65 000	Jung et al. (2010)	36 687	Döll et al. (2003)
65 500	Oki and Kanae (2006)	37 288	Dai and Trenberth (2002)
66 000	Sterling et al. (2012)	38 587	Baumgartner and Reichel (1975)
71 000	Baumgartner and Reichel (1975)	38 605	Widén-Nilsson et al. (2007)
72 000	Korzun (1978)	39 307	Fekete et al. (2002)
75 981 ^b	Mueller et al. (2011)	39 414	Döll and Fiedler (2008)
60 000–85 000	Haddeland et al. (2011)	44 560	Korzun (1978)
		45 500	Oki and Kanae (2006)
		42 000–66 000	Haddeland et al. (2011)
70 576	STANDARD	40 458	STANDARD

^a 1.35 mm d^{-1} based on a land area of $130.922 \times 10^6 \text{ km}^2$ ^b 1.59 mm d^{-1} based on a land area of $130.922 \times 10^6 \text{ km}^2$ (value taken from Mueller et al., 2013, as no area is given in Mueller et al., 2011).

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Table 7. Rank of model variants where global land area (except Greenland and Antarctica) is affected most based on a threshold which represents the 10th percentile of averaged (1971–2000) global grid cell values for AET and discharge.

rank	variant	% of area affected by changes above 10th percentile	
		AET	discharge
1	NoCal	60.5	13.5
2	CLIMATE	45.5	3.2
3	LANDCOVER	24.2	1.2
4	STRUCTURE	13.6	1.1
5	NoUse	0.9	0.03

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Table A1. Parameters of the leaf area index model.

no.	land cover type	L_{\max} [-]	fraction of deciduous plants $f_{d,lc}$	L reduction factor for evergreen plants $C_{e,lc}$	initial days to start/end with growing season [d]
1	Evergreen needleleaf forest	4.02 ^a	0	1	1
2	Evergreen broadleaf forest	4.78 ^b	0	0.8	1
3	Deciduous needleleaf forest	4.63	1	0.8	10
4	Deciduous broadleaf forest	4.49 ^c	1	0.8	10
5	Mixed forest	4.34 ^d	0.25	0.8	10
6	Closed shrubland	2.08	0.5	0.8	10
7	Open shrubland	1.88	0.5	0.8	10
8	Woody savanna	2.08	0.5	0.3	10
9	Savanna	1.71	0.5	0.5	10
10	Grassland	1.71	0	0.5	10
11	Permanent wetland	6.34	0	0	10
12	Cropland	3.62	0	0.1	10
13	Cropland/natural vegetation mosaic	3.62	0.5	0.5	10
14	Snow and ice	0	0	0	0
15	Bare ground	1.31	0	1	10

^a L_{\max} is assumed to be the mean value of land cover classes of Scurlock et al. (2001) TeENL and BoENL,

^b only value for TrEBL and not TeEBL (Scurlock et al., 2001) as in WaterGAP this class is mainly in the tropics,

^c mean value from TeDBL and TrDBL (Scurlock et al., 2001),

^d mean value of all forest classes. Fraction of deciduous plants and L reduction factor for evergreen plants based on IMAGE (Alcamo et al., 1998), initial days to start/end with growing season are estimated.

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Table A2. Attributes for IGBP land cover classes used in WaterGAP 2.2 for all model variants, compiled from various literature sources. Water has an albedo of 0.08, snow 0.6.

no.	land cover type	rooting depth ^a [m]	albedo ^a [-]	snow albedo [-]	emissivity ^b [-]	degree-day factor D_F^c [mm d ⁻¹ °C ⁻¹]
1	Evergreen needleleaf forest	2	0.11	0.278	0.9956	1.5
2	Evergreen broadleaf forest	4	0.07	0.3	0.9956	3
3	Deciduous needleleaf forest	2	0.13	0.406	0.99	1.5
4	Deciduous broadleaf forest	2	0.13	0.558	0.99	3
5	Mixed forest	2	0.12	0.406	0.9928	2
6	Closed shrubland	1	0.13	0.7	0.9837	3
7	Open shrubland	0.5	0.2	0.7	0.9541	4
8	Woody savanna	1.5	0.2	0.558	0.9932	4
9	Savanna	1.5	0.3	0.7	0.9932	4
10	Grassland	1	0.25	0.7	0.9932	5
11	Permanent wetland	1	0.15	0.2	0.992	4
12	Cropland	1	0.23	0.376	0.9813	4
13	Cropland/natural vegetation mosaic	1	0.18	0.3	0.983	4
14	Snow and ice	1	0.6	0.7	0.9999	6
15	Bare ground	0.1	0.35	0.7	0.9412	6

^a adapted from the IMAGE model (Alcamo et al., 1998)^b (Wilber et al., 1999)^c (Maniak, 1997; WMO, 1994)

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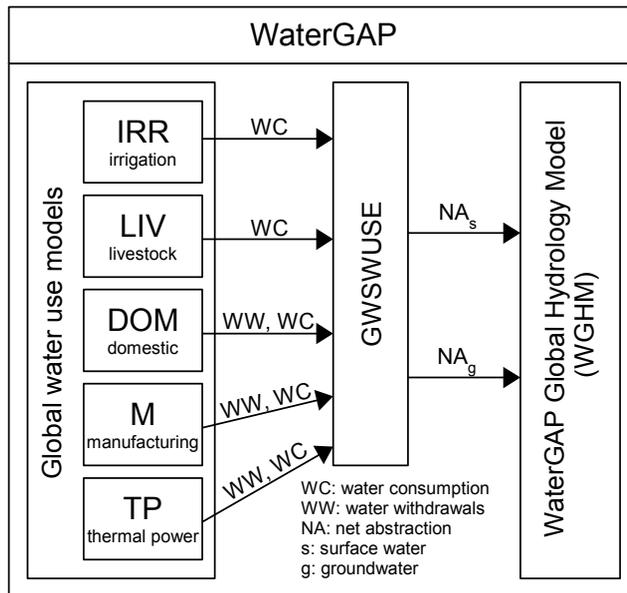


Fig. 1. Schematic of WaterGAP 2.2. The output of five water use models is translated into net abstractions from groundwater NA_g and surface water NA_s by the submodel GWSWUSE, which allows computing the impact of human water use on water flows and storages by WGHM. For details see Döll et al. (2012).

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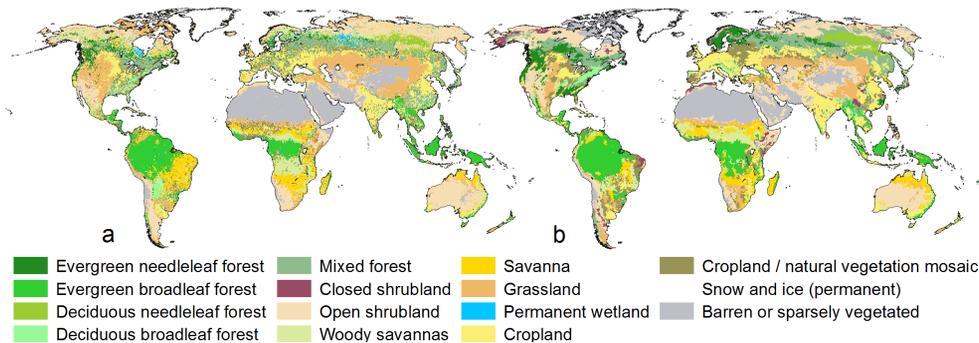


Fig. 2. Land cover maps with a spatial resolution of 0.5° used as WaterGAP input based on MODIS observations for the year 2004 (variant STANDARD) **(a)**, and land cover derived from USGS GLCC but CORINE for Europe reflecting land cover distribution around the year 2000 (variant LANDCOVER) **(b)**.

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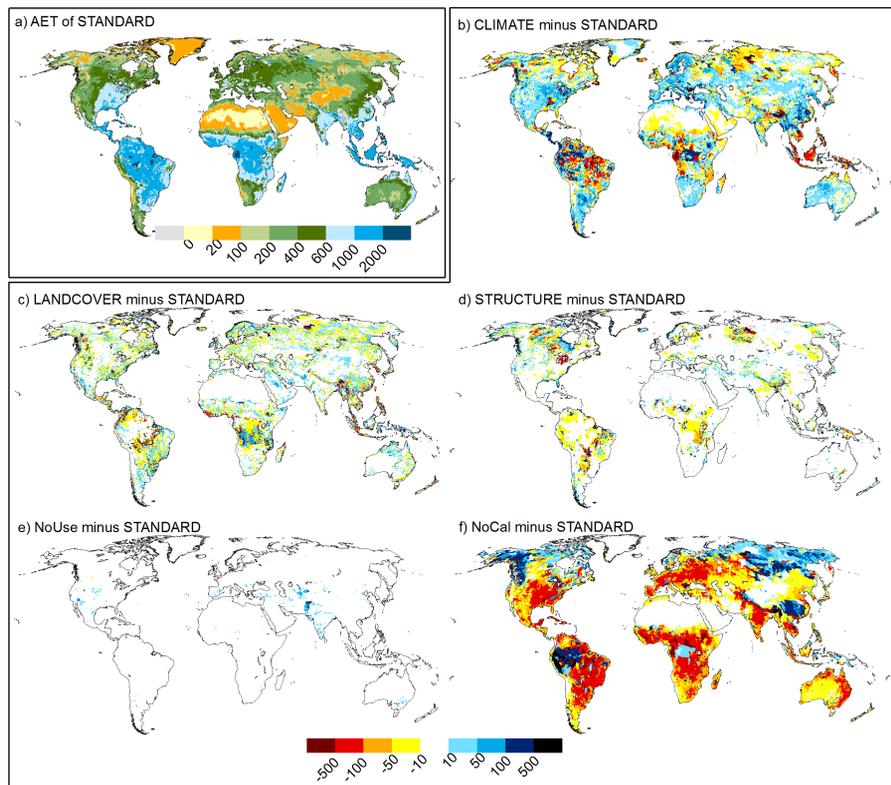


Fig. 3. Actual evapotranspiration AET for STANDARD (mean value 1971–2000, in mm yr^{-1}) (a) and absolute differences between the model variants and STANDARD (b–f).

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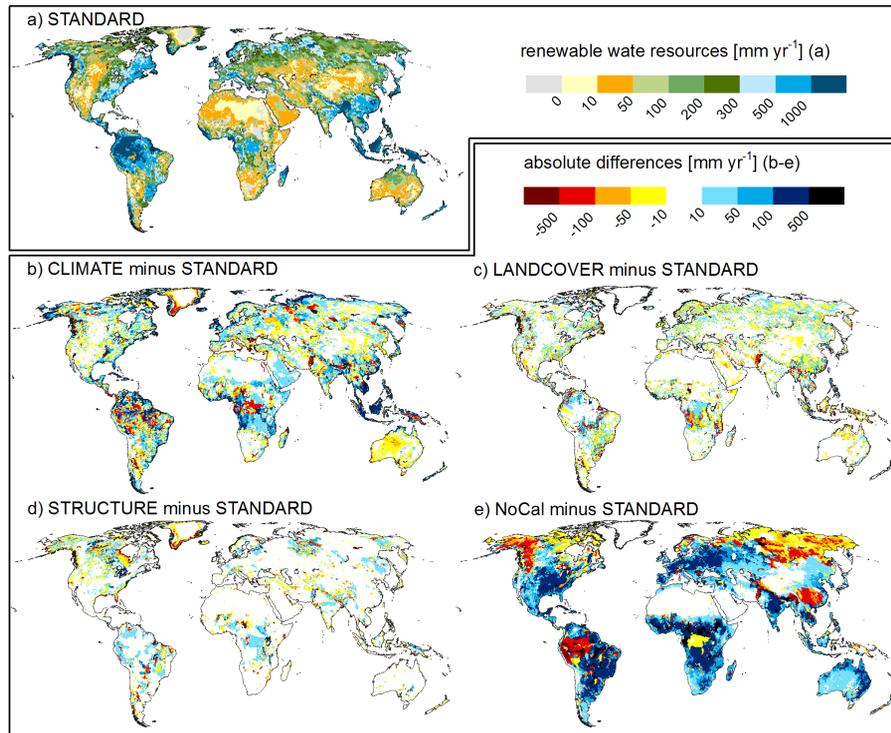


Fig. 4. Renewable water resources (mean annual runoff from each cell if water use is neglected) calculated by WaterGAP 2.2 STANDARD variant (a) and absolute differences to other variants (b–e).

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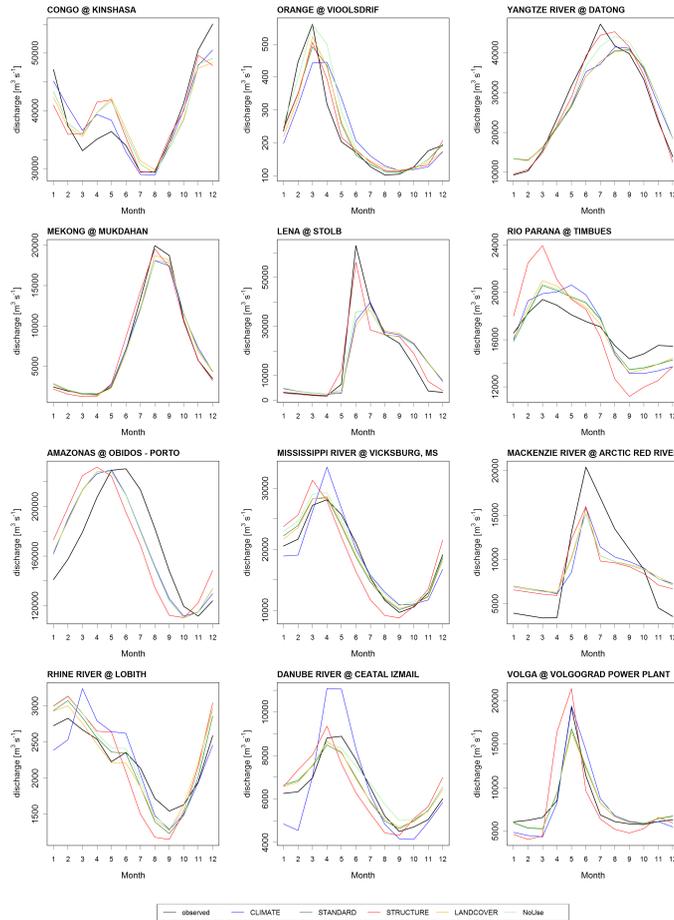


Fig. 5. Discharge seasonality for selected basins and the calibrated model variants.

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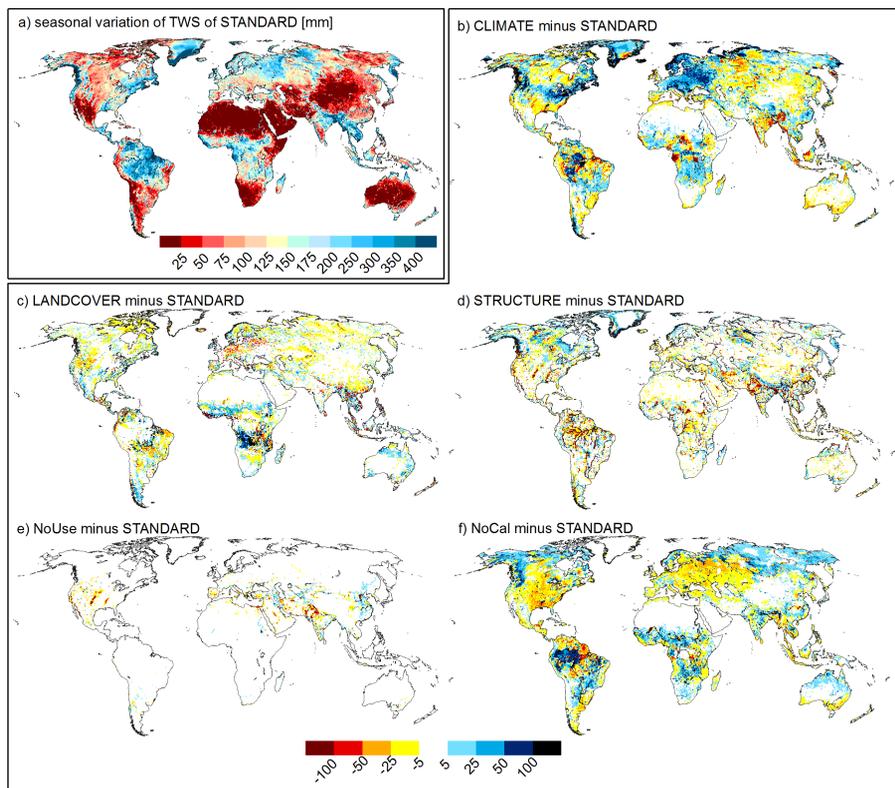


Fig. 7. Seasonal variation of total water storage (TWS) for STANDARD (a) and as absolute difference maps [mm] to all other model variants (b–f).

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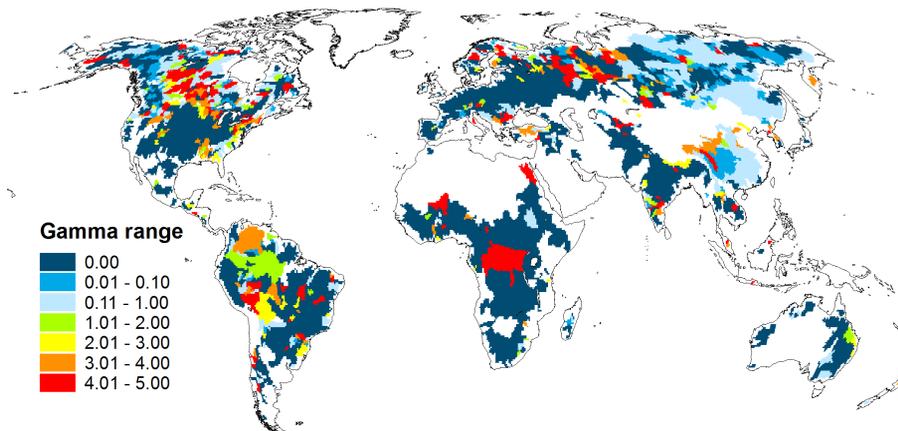


Fig. 8. Range of calibration parameter γ through all four calibrated model variants (calculated as $\gamma_{\max} - \gamma_{\min}$) showing the general sensitivity to input data and model structure. White colors indicate uncalibrated regions.

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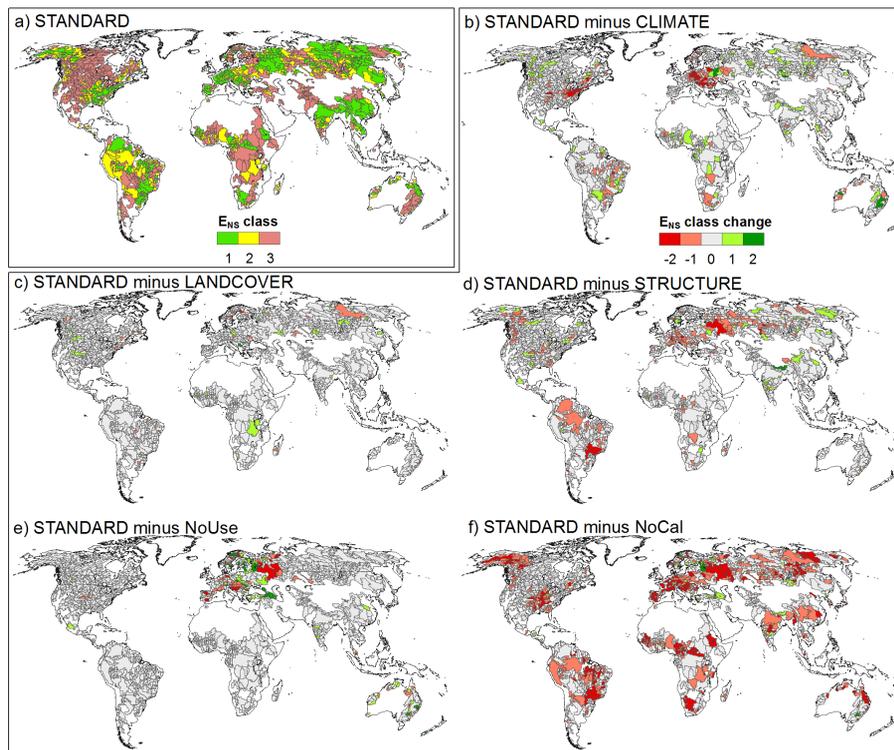


Fig. 9. Spatial distribution of Nash–Sutcliffe efficiency E_{NS} classes (from Table 4, 1: $E_{NS} > 0.7$, 2: $0.5 < E_{NS} < 0.7$, 3: $E_{NS} < 0.5$) for STANDARD (a), and differences of model variants (calculated as STANDARD E_{NS} class minus that of the model variant) (b–f). Red colors indicate a decrease, green an increasing E_{NS} when using the model variant compared to STANDARD.

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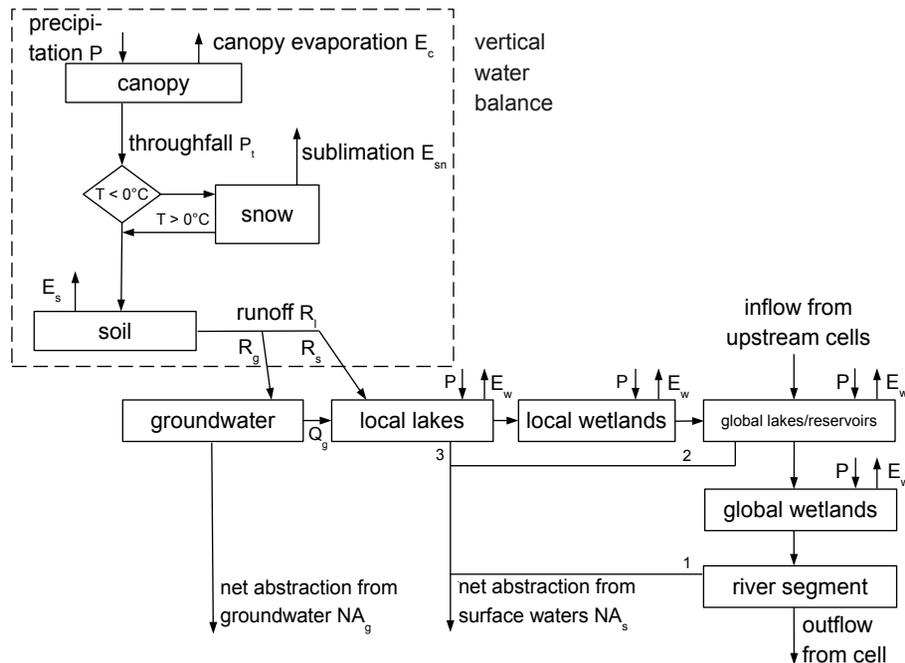


Fig. A1. Schematic structure of the water fluxes and storages as computed by WaterGAP Global Hydrology Model (WGHM) within each 0.5° grid cell. Boxes represent water storage compartments, arrows water fluxes (inflows, outflows). Numbers at net abstraction from surface waters (NA_s) are the order from which storage water is abstracted until demand is satisfied.

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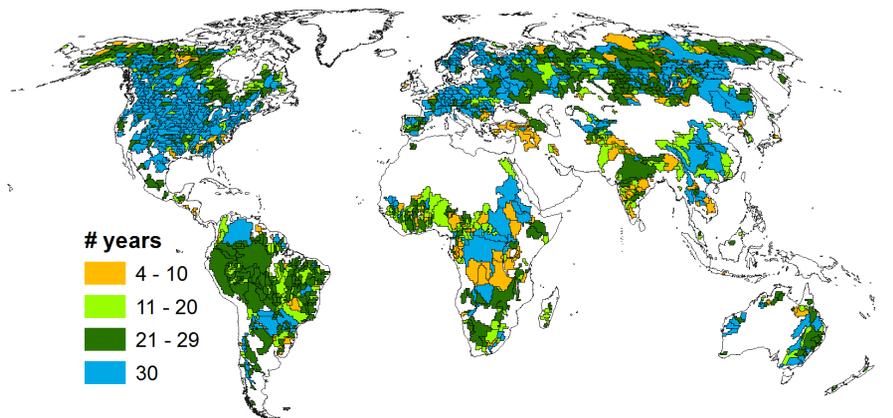


Fig. B1. Calibration basins of WaterGAP 2.2 with number of years with discharge observations used for calibration.

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